# Reversible Sorption of Nitrogen Oxides in Mn–Zr Oxide

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mixed metal oxide systems in the presence and absence of and  $SO_2$  (12, 13). This difficulty is to some extent related gaseous  $O_2$ . At 200–300°C Mn–Zr oxide exhibited the highest to the very dilute concentration of NO i gaseous  $O_2$ . At 200–300°C Mn–Zr oxide exhibited the highest to the very dilute concentration of NO<sub>x</sub> in exhaust gases.<br>NO removal among the oxides investigated and it was facili-<br>tated in the presence of  $O_2$ . The rem where the solution of oxidized NO species on of in the solid and stored<br>via formation of nitrate. Since the sorption and desorption was<br>reversible, the oxide sorbent was regenerated on heating the<br>sample at 400°C after NO **were also eliminated from the solid by reduction with**  $C_3H_8$  **or** selective reduction of NO<sub>x</sub> by a 3-way catalyst and decom-<br> $C_3H_3OH$  at 300°C. The amorphous phase of the Mn-Zr oxide position of NO (14), is operated **C2H5OH at 300**8**C. The amorphous phase of the Mn–Zr oxide** position of NO (14), is operated when NO is released from which is formed after heating this oxide at 450°C was extremely **effective for NO removal due to good mixing between Mn** coexisting large amounts of oxygen or water. A similar **and Zr oxides. The surface area of the oxide was a maximum concept has been commercialized in the**  $NO<sub>x</sub>$ **-storage re-<br>and the amorphous phase was dominant for the oxide with duction catalyst proposed by the research gr** 

for solving acid rain and other air pollution problems in the atmosphere. For large power stations, selective catalytic ible NO absorbent or adsorbent in the catalytic conversion reduction (SCR) of NO<sub>x</sub> with ammonia is now well-estab-of NO<sub>x</sub>. Another possibility is to treat the concentrated lished technology (1). However, for relatively small scale  $NO<sub>x</sub>$  by a wet process. The present investigation aims at combustors for cogeneration systems or on-site generators, removal of  $NO<sub>x</sub>$  from dilute gaseous mixture to form conthe high capital cost for the SCR with ammonia and the centrated nitrogen oxides by sorption–desorption cycles possibility of a slip of ammonia become problems. Three- of a solid oxide. Nitrogen oxides are stored in the so possibility of a slip of ammonia become problems. Threeway catalysts have been widely used for cleaning automo- until the metal oxide is converted into the corresponding tive exhaust when gasoline engines are operated near the nitrate salt, whereupon stored NO*<sup>x</sup>* is again released into stoichiometric air/fuel ratio (2). However, a new and inno- the gas phase in a concentrated form on heating the solid. vative process has been requested for cleaning exhaust A related phenomenon has already been reported for Y– gases from diesel-powered vehicles and those with lean Ba–Cu–O (16), Ba–Cu–O (17), and Y–Sr–Co–O (18) sysfuel engines. These applications require a catalyst to be tems. These oxides have been reported to be active for operated under oxidative conditions. Selective reduction removal of a large amount of  $NO<sub>x</sub>$  by absorption, but of NO*<sup>x</sup>* with hydrocarbons has been actively investigated deactivation of the absorbents is expected because of the recently by many researchers  $(3, 4)$ . Cu-exchanged ZSM- formation of surface carbonate in the presence of  $CO<sub>2</sub>$ . 5 (5, 6), metallosilicates (7), alumina (8), and metal-ex- The rare earth and alkaline earth metal cations in these changed zeolites (9–11) have been reported to be active oxide systems serve as strong absorption sites for  $CO<sub>2</sub>$  due for this reaction. However, SCR with hydrocarbons has to their strong basicity. We have tried to develop materials

not been commercialized up to now due to problems of **Removal of NO (900 ppm) was investigated using several** catalyst deactivation by the coexisting gases, such as H<sub>2</sub>O mixed metal oxide systems in the presence and absence of and SO<sub>2</sub> (12, 13). This difficulty is to some and the amorphous phase was dominant for the oxide with<br>Mn/Zr = 1. The removal of NO was not influenced by coexisting<br>CO<sub>2</sub> and was promoted by the presence of H<sub>2</sub>O.  $\otimes$  1996 Academic<br>Press, Inc.<br>Press, Inc.<br>Press, Inc form during the cruising of an automobile in lean-burn **CONDUCTION CONDUCTION CONDUCTION** by the 3-way catalyst when the atmosphere is changed Removal of NO<sub>x</sub> from exhaust gases is a crucial problem intermittently to the stoichiometric air/fuel ratio. This ap-<br>*x* r solving acid rain and other air pollution problems in the plication suggests the importance of i

for  $NO<sub>x</sub>$  absorption which do not contain rare earth and TABLE 1 alkaline earth components. In this paper, novel sorption–<br>desorption properties of Mn–Zr oxides are reported for **Containing Mn and/or Zr**<br>**Containing Mn and/or Zr** application to NO<sub>x</sub> removal.

## **EXPERIMENTAL**

Mixed oxide samples were prepared by coprecipitation from a solution of the corresponding nitrate mixtures. Various mixed oxide samples were calcined at 450°C prior to the NO removal experiments. For the preparation of Mn–Zr oxides, the calculated amounts of  $Mn(NO<sub>3</sub>)<sub>3</sub>$  and  $ZrO(NO<sub>3</sub>)<sub>2</sub>$  (Kishida Chemical, Guaranteed reagent grade) were dissolved in water. After adding aqueous ammonia to the nitrate solution, the solution with the precipitate was evaporated to dryness and then heated at  $450^{\circ}$ C for 6 h. The powder thus obtained was sieved to  $10-20$ mesh.  $ZrQ_2 \cdot MoQ_3$  0.3 2.4 mesh.<br>Sorption and desorption experiments of NO were car-

ried out in a flow system. Nitrogen monoxide was diluted *Note.* Calcination temperature 450°C. 0.1 in He and sometimes mixed with  $O_2$  before admission to vol.% NO, 0 or 10% O<sub>2</sub>, He balance. Reaction a quartz tube reactor. The untake and desorption charaction a quartz tube reactor. The uptake and desorption charac-<br>teristics of NO<sub>*x*</sub> were tested in the temperature range from  $a$  NO removal after 30 min of use. 200 to 400°C at a contact time of 0.6–1.0 g s cm<sup>-3</sup> using 1 g of the mixed oxide sample. When required,  $CO<sub>2</sub>$ ,  $H<sub>2</sub>O$ , and/or  $C_3H_8$  was added to the gaseous mixture. The con-<br>centration of NO was analyzed with a chemical lumines-<br>tailed NO removal characteristics and mechanisms were centration of  $NO_x$  was analyzed with a chemical lumines-<br>cence type  $NO$  meter (Shimadzu  $NOA305$ ) The effluent accordingly investigated for the Mn–Zr system. cence type NO<sub>x</sub> meter (Shimadzu, NOA305). The effluent gas was generally passed through a carbon reactor to re-<br>duce NO<sub>2</sub> for analysis of the sum of NO and NO<sub>2</sub> concen-<br> $NO-氨orptive$  *Properties for Mn–Zr Oxide* trations before it was supplied to the  $NO_x$  meter. The The time course of the NO removal for the Mn–Zr ox-<br>gaseous composition of the inlet and outlet mixture was ide is shown in Fig. 1. Nitrogen monoxide (900 ppm)

### *Removal of NO by Mixed Metal Oxides*

Removal of NO was investigated both in the presence and in the absence of  $O_2$  at 200°C in NO (900 ppm) for several oxide materials containing Mn and/or Zr as listed in Table 1. Only the oxides which exhibited a higher removal of NO in the presence of  $O_2$  than in the absence of  $O<sub>2</sub>$  after 30 min from the start of the reaction are listed in the table since the present investigation aims at NO removal in air. The mechanism of NO removal was not taken into account at this stage; some oxides are active for decomposition of NO and some are effective in absorption or adsorption. In the series of oxides studied, the Mn–Fe, Mn–Cr, Mn–Zr, and Mn–Cu systems exhibited a relatively high removal of NO. The Mn–Zr oxide exhibited the high-<br>**FIG. 1.** Time course of NO removal by Mn–Zr oxide (Mn/Zr = 1). est removal of NO both with and without  $O_2$ . Nitrogen  $900$  ppm NO,  $10\%$   $O_2$ , He balance. Reaction temperature 200°C. monoxide was scarcely detected in the outlet gas after  $W/F = 1 g s cm^{-3}$ .

Oxide	NO removal <sup><i>a</i></sup> $(\%)$	
	$0\%$ O <sub>2</sub>	$10\%$ O <sub>2</sub>
$MnOx \cdot Al_2O_3$	9.9	14.8
$MnOx \cdot Cr_2O_3$	0	2.2
$MnOx \cdot CuO$	0	10.2
$MnOx \cdot Fe_2O_3$	11.9	37.4
$MnOx \cdot MoO_3$	5.9	15.2
$MnOx \cdot TiO_2$	0	25.0
$MnOx \cdot ZrO_2$	100.0	100.0
$ZrO_2 \cdot Al_2O_3$	0	8.7
$ZrO_2 \cdot Cr_2O_3$	$\Omega$	14.3
$ZrO_2 \cdot CuO$	2.7	27.0
$ZrO_2 \cdot Fe_2O_3$	0.4	8.7
$ZrO_2 \cdot MoO_3$	0.3	2.4
$ZrO_2 \cdot TiO_2$	7.6	17.4

100

gaseous composition of the inlet and outlet mixture was ide is shown in Fig. 1. Nitrogen monoxide (900 ppm) also analyzed by gas chromatography. The phase in the and oxygen (10%) was supplied to the Mn-Zr oxide also analyzed by gas chromatography. The phase in the and oxygen (10%) was supplied to the Mn–Zr oxide sample was identified with an X-ray diffractometer (Ri-  $(Mn/Zr = 1)$  at 200°C. Most of the NO was removed during sample was identified with an X-ray diffractometer (Ri-  $(Mn/Zr = 1)$  at 200°C. Most of the NO was removed during gaku, RINT-1400). the initial 60 min, then the concentration of NO at the outlet gradually increased with an elapse of time. The **RESULTS AND DISCUSSION** sample no longer showed a removal property after 6 h

80 NO removal / % 60 40 20  $0\frac{L}{0}$  $10$ 2 Time /h



from the start of the experiment. The total amount of NO  $ZrO_2$  rich composition. In every sample except the Zr-rich<br>removed in 6 h of operation was 14.7 cm<sup>3</sup> (STP)/g which sample (Mn/Zr = 1/9), the diffraction lines ar formation of oxidized NO species as discussed later. The *Sorption–Desorption Behavior of Mn–Zr Oxide* NO uptake measured by the gravimetric analysis was some-<br>what smaller than that from the tubular reactor by gas-<br>phase analysis. The difference in storage capacity for the<br>phase analysis. The difference in storage capacity two measurements is probably due to gas supply in the macropores. The NO-containing gas was forced to pass **TABLE 2** through a solid bed for the tubular reactor, whereas the gas diffused in a powder agglomerate in the case of the gravimetric analysis. As we have reported previously, absorption of NO in Ba–Cu oxide or Ba–Cu–Mn oxide proceeds via oxidation of NO and the formation of nitrate in the solid bulk (17). The present NO uptake for Mn–Zr oxide resembles the previous system and is expected to proceed also by the formation of nitrate, as explained later.

The sorptive properties of the Mn–Zr oxides for NO were investigated using samples with different Mn/Zr ratios (Table 2). The removal curves were basically similar to that shown in Fig. 1. The NO removal was 100% for every Mn–Zr oxide at the start of the NO supply and then *Note.* Calcination temperature 450°C. 900 ppm NO, 10%  $O_2$ , He balgradually decreased after 10 to 60 min of operation. The ance Reaction temperature 200°C, W/F = 1 g s cm<sup>-3</sup>.

start of the decrease and therefore the sorption capacity was quite dependent on the Mn/Zr ratio of the oxide. The amount of stored NO was largest for the sample with Mn/  $Zr = 1$  but decreased on either decreasing or increasing the Mn/Zr ratio. The addition of noble metals was not effective in promoting the absorption capacity and absorption rate, as can be seen in Table 2.

The X-ray diffraction patterns of Mn–Zr oxides with different Mn/Zr ratios after calcination at  $450^{\circ}$ C are shown in Fig. 3. At the Mn-rich compositions, very weak diffraction lines of  $MnO<sub>2</sub>$  were detected in the pattern. Diffraction lines from  $ZrO<sub>2</sub>$  or reaction products was not observed at  $Mn/Zr > 1/5$ . The sample at  $Mn/Zr = 5$  consisted of  $Mn_2O_3$  and  $Mn_5O_7$  phases. Only a weak diffraction line of **FIG. 2.** Weight change of Mn–Zr oxide (Mn/Zr = 1) during NO  $Mn_2O_3$  was observed for the sample with Mn/Zr = 1 in removal. 850 ppm NO, 10% O<sub>2</sub>, He balance. Reaction temperature 200 °C. addition to a broad halo from an amorphous phase. For the sample with  $Mn/Zr = 1/5$ , diffraction lines from the  $ZrO<sub>2</sub>$  phase appeared and these were intensified at the







The since neither sorption nor desorp-<br>300°C. The outlet concentration of NO was almost zero<br>as shown in the figure. The concentration of NO in the<br>gas-phase sharply increased from 320°C, then exceeded<br>lower than the inle gas-phase sharply increased from  $320 \text{ C}$ , then exceeded lower than the inlet value in the temperature range from the value of the feed gas. This behavior confirms that the  $300 \text{ to } 600^{\circ}\text{C}$ . The concentration of N removal of NO for this system proceeds by sorption of<br>
NO or oxidized NO<sub>x</sub> species in the solid and the desorption<br>
occurs at elevated temperatures. The outlet concentration<br>
of NO sharply decreased after passing through



ppm NO + 10% O<sub>2</sub>, ( $\triangle$ ) 900 ppm NO + 1500 ppm C<sub>3</sub>H<sub>8</sub>, He balance.  $W/F = 1 g s cm^{-3}$ . (---) initial concentration of NO.

The temperature-programmed sorption–desorption curves in other atmospheres in which the sample was heated in NO (900 ppm) + He and in NO (900 ppm) +  $C_3H_8$  (1500 ppm) + He are also shown in Fig. 4. In every case, the removal of NO proceeded at low temperatures; thus the NO concentration in the outlet gas was almost zero. With increasing temperature, the NO concentration rose abruptly, indicating desorption of NO. The temperature of the transition from uptake to desorption was quite sensitive to the atmosphere. The desorption started from the lowest temperature in the reducing atmosphere with  $C_3H_8$ , whereas the desorption started from the highest temperature in the  $O_2$ -containing atmosphere. It is obvious **FIG. 3.** XRD pattern of Mn/Zr oxides with different Mn/Zr ratios that the start of the desorption began at lower temperatures after heating at 450°C. Values in parentheses are surface areas. (O) as the experiment was oper after heating at 450°C. Values in parentheses are surface areas. ( $\circ$ ) as the experiment was operated in more reducing atmo-<br>Mn<sub>2</sub>O<sub>3</sub>, ( $\Box$ ) MnO<sub>2</sub>, ( $\bullet$ ) Mn<sub>5</sub>O<sub>7</sub>, ( $\Box$ ) ZrO<sub>2</sub>. sphere. In other words, the oxidizin the sorbed species. It is expected that the uptake of NO stant rate (Fig. 4). The Mn-Zr oxide sample was heated<br>at 100°C and exposed to the atmosphere of flowing NO<br>(900 ppm) with O<sub>2</sub> (10%) and then heated at the constant<br>rate of 10°C/min. The removal of NO proceeded below<br>rat

> $(Mn/Zr = 1)$  at 200°C (Fig. 5). Nitrogen monoxide was completely removed from the gas phase for the initial 60 min. Then, the sample was exposed to He and heated within 30 min to 400°C. The sample was kept at  $400^{\circ}$ C in He until the stored  $NO<sub>x</sub>$  was completely eliminated. The amount of NO uptake (0.047 mol/mol-Zr) almost agreed with that of desorption (0.044 mol/mol-Zr), indicating that the sorption and desorption is almost reversible.

The gas-phase analysis indicated that only nitrogen monoxide was released in the desorption step in most cases. There may be several possible desorption species, but nitrous oxide  $(N_2O)$  was scarcely observed by gas chromatography during the desorption experiment. Of two other desorption species, i.e.,  $NO$  and  $NO<sub>2</sub>$ , only nitrogen monoxide is generally released due to its stability in the desorp-FIG. 4. Sorption-desorption curve of NO for Mn-Zr oxide<br>
(Mn/Zr = 1) during programmed heating in different atmospheres<br>
after sample was heated at 450°C for 6 h. (O) 900 ppm NO, ( $\square$ ) 900<br>
ppm NO + 10% O<sub>2</sub> ( $\triangle$ ) 900 p

The temperature-programmed sorption–desorption ex-



balance.  $T = 200^{\circ}$ C (60 min), W/F = 0.6 g s cm<sup>-3</sup>. Desorption condition: catalytic converter.  $200^{\circ}\text{C} \rightarrow 400^{\circ}\text{C}$  and at  $400^{\circ}\text{C}$  in He.

addition to the thermal desorption, the reduction treat-<br>ment is considered as an alternative method to recover<br>the a surface or a bulk reaction. The overall amount of<br>the oxide solid ofter the NO uptake The NO corption w the oxide solid after the NO uptake. The NO sorption was<br>carried out in NO (900 ppm) for 10 min at 300°C in the mol-NO/mol-Zr) is too large to be explained by a surface<br>monolayer reaction. However, the extraordinary large presence of  $O_2$  (10%), then the atmosphere was switched<br>to the reducing atmosphere of  $C_3H_8$  or  $C_2H_5OH$  at the face area and the amorphous formation at this composition<br>still suggest the possibility of surface adsor same temperature, as shown in Fig. 6. The removal of NO<br>was 100% for 10 min. The desorption peak which appeared<br>after switching the atmosphere was more sharp in the  $C_3H_8$ <br>atmosphere than in  $C_2H_5OH$ . The amount of rem NO was 8.1  $\times$  10<sup>-3</sup> mol/mol-Zr, whereas the desorption



balance.  $T = 300^{\circ}\text{C}$  (10 min). W/F = 0.6 g s cm<sup>-3</sup>. Desorption condition:  $W/F = 0.6$  g s cm<sup>-3</sup>.

amounted to 7.0  $\times$  10<sup>-3</sup> mol-NO/mol-Zr for C<sub>3</sub>H<sub>8</sub> and  $4.4 \times 10^{-3}$  mol-NO/mol-Zr for C<sub>2</sub>H<sub>5</sub>OH, namely only 86 and 55% of NO was reversibly desorbed, respectively. The remaining parts were reduced with  $C_3H_8$  or  $C_2H_5OH$  and therefore the formation of  $N_2$  was detected in the outlet gas by gas chromatography. It is noted from the comparison of Figs. 5 and 6 that the desorption proceeds more rapidly in the reducing atmosphere than in the case of He. The desorption was completed within 25 min at  $300^{\circ}$ C by using  $C_3H_8$  or  $C_2H_5OH$ , even though the temperature was low. For the recovery of the oxide, the treatment in the controlled reducing atmosphere is more effective and easier than the treatment of the sample at high temperatures. FIG. 5. Amounts of NO sorbed in Mn-Zr oxide (Mn/Zr = 1) at<br>200°C and desorbed at 400°C. Hatched areas indicate amounts of NO<br>sorbed and desorbed. Sorption condition: 900 ppm NO, 10% O<sub>2</sub>, He<br>balance.  $T = 200$ °C (60 min).

### *Mechanism of NO Uptake*

periments in different atmospheres suggest that the reduc-<br>tion of stored species facilitates the desorption of NO<sub>x</sub>. In<br>example of nitrate-like species on or in the solid. It is not<br>edition to the thermal decoration the at 1380 cm<sup>-1</sup> which is attributed to the  $NO_3^-$  species on or in the solid (19). The behavior of NO removal resembles that observed in the alkaline earth or rare earth oxide containing systems in which bulk reaction is reported to be operative (17, 18).

The NO uptake for the Mn–Zr oxide is expected to be initiated by the catalytic oxidation of NO to  $NO<sub>2</sub>$ . The oxidation activity for NO to  $NO<sub>2</sub>$  and  $NO<sub>3</sub>$  is therefore important for this type of NO removal system. The oxidation of NO to  $NO<sub>3</sub>$  ion appears to proceed via the formation of  $NO<sub>2</sub>$  on the solid surface. The catalytic oxidation of NO was tested in the flow system using pure Mn oxide. The temperature dependence of the oxidation activity is shown in Fig. 7. The equilibrium conversion of the reaction  $NO + 1/2 O_2 \rightleftarrows NO_2$  in 10%  $O_2$  is shown by the solid line in the figure. The conversion of NO increased initially as FIG. 6. Amounts of NO sorbed in Mn-Zr oxide and desorbed in<br>reducing atmospheres. Sorption condition: 900 ppm NO, 10% O<sub>2</sub>, He<br>high activity of Mn oxide for NO oxidation to NO<sub>2</sub>. 1500 ppm  $C_3H_8$  or 1500 ppm  $C_3H_3OH$ , He balance,  $T = 300^{\circ}$ C and Thus, the Mn species in Mn–Zr oxide serves as an oxida-. tion catalyst in the presence of oxygen. The NO absorption



. (s) NO and (h) 900 ppm NO, 10% O2, He balance. W/F <sup>5</sup> 1 g s cm<sup>2</sup><sup>3</sup> . ( ) equilib- NO2 removal. rium conversion.

also proceeded in the absence of gaseous  $O_2$ , as can be<br>seen in the low temperature region of Fig. 4. The valence<br>of Mn oxide is known to be very sensitive to the change<br>in atmosphere and temperature in the present reac

first investigated because the exhaust gas generally conpresent system, the removal of NO is suggested to be initiated by the oxidation of NO to  $NO<sub>2</sub>$  on the surface. Therefore, the removal of  $NO<sub>2</sub>$  is expected to proceed easily on the Mn–Zr oxide, which is confirmed from the comparison of curves for NO and  $NO<sub>2</sub>$  removal in Fig. 8. The curve for  $NO<sub>2</sub>$  removal was almost similar to the case of NO removal. The overall amount of  $NO<sub>2</sub>$  uptake (0.107) mol/mol-Zr) was slightly larger than that of NO (0.082 mol/mol-Zr).

The behavior of NO removal was also tested in the presence of  $CO<sub>2</sub>$  and  $H<sub>2</sub>O$ , since these gases are always present in exhaust gas (Fig. 9). Many solid oxide absorbents which operate at high temperatures deteriorate in the presence of these gases. Our previous study on Cu–Ba or Cu–Mn systems showed no obvious deterioration with  $CO<sub>2</sub>$ within a few days, but long operation in CO2 atmosphere **FIG. 9.** Time course of NO removal with or without CO2 or steam gradually deactivated the absorption capacity due to strong for Mn–Zr oxide (Mn/Zr <sup>5</sup> 1). (s) 900 ppm NO, 10% O2; (h) 900 ppm bonding between  $CO<sub>2</sub>$  and alkaline earth or rare earth cations (17). However, in the present Mn–Zr system, the  $T = 200^{\circ}$ C, W/F = 1 g s cm<sup>-3</sup>.



**FIG. 7.** Activity of Mn oxide catalyst for oxidation of NO to NO<sub>2</sub>. **FIG. 8.** NO and NO<sub>2</sub> removal curves for Mn–Zr oxide (Mn/Zr = 0.000 nm NO 10% O. He balance. *T* = 2008C, W/F = 1 g s cm<sup>-3</sup>. (O) NO and ( $\square$ )

conditions. Therefore, oxygen species present in the Million The behavior in the presence of H<sub>2</sub>O is also shown in oxide lattice before the NO uptake is used for oxidation of NO and subsequent sorption. Thus the sorption *Effect of Coexisting Gases on Removal of NO* start of the reaction. The overall amount of uptake was *Effect of Coexisting Gases on Removal of NO* showiously enhanced with H<sub>2</sub>O. It is important that the The effect of coexisting gas is especially important be- sorption is promoted in the presence of  $H_2O$  because comcause NO absorption is desired to be applied for cleaning bustion exhaust always contains moisture. The adsorbed exhaust gases from combustors. The removal of  $NO<sub>2</sub>$  was or absorbed water molecules appear to stabilize sorbed first investigated because the exhaust gas generally con-<br>intrate ions through hydration and to facilitat tains a nonequilibrated mixture of NO and NO2. In the reaction by promotion of diffusion. However, a detailed



investigation is necessary for the exact understanding of **REFERENCES** the promotion effect of  $H_2O$ .

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The sorption–desorption behavior of solids has not been<br>
investigated satisfactorily so far because research on  $NO_x$ <br>  $_{L15(1991)}$ <br>  $_{L15(1991)}$ <br>  $_{L15(1991)}$ removal has been mainly focused on the development of 4. Misono, M., and Kondo, K., *Chem. Lett.,* 1001 (1991). the catalytic process. The present system proposed a possi- 5. Iwamoto, M., Yahiro, H., Shundo, Y., Yu, O., and Mizuno, N., bility of the combination of a catalytic process with the Shokubai, *Catalysts* **32**, 430 (1990).<br>
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